CAS # 375-95-1	Perfluorononanoic acid
	Synonym ¹ s: EINECS 206-801-3; UNII-5830Z6S63M; HSDB 8040
	RTECS # ² : RA6901000
	EINECS # ³ : 206-801-3
FFFFFFF	Molecular Weight4: 464.0709
F	Molecular Formula ⁵ : C9-H-F17-O2
F F F F F F F OH	
PHYSICAL CHARACTERISTICS	
Primary Use	Long-chain perfluoroalkanecarboxylic acids and their salts are surface- active chemicals (surfactants), which greatly reduce the surface tension (surface energy) of water, aqueous solutions, and organic liquids even at low concentrations. These acids (C_6 - C_{12}) and derivatives are used as
	wetting, dispersing, emulsifying, and foaming agents. /Long-chain
	perfluoroalkanecarboxylic acids/6
	Breakdown product of stain- and grease-proof coatings on food
	packaging, couches, carpets. ⁷
Physical state, odor at room	White, solid ⁸
temperature & pressure	
Melting point; Boiling point	MP/range: 68-73°C (154-163°F) ⁹ ; Initial BP and Boiling range = 218°C (424°F) at 987 hPa (740 mmHg) ¹⁰
Solubility	No data available
Specific Gravity	No data available
SAFETY/PHYSICAL HAZARDS	
Vapor Pressure	8.3X10 ⁻² mm Hg at 25 deg C (est) ¹¹
Flammability	No data available
Flashpoint	No data available
Flammability Rating	No data available
Auto Ignition Point	No data available
Combustion products	Hazardous decomposition products formed under fire conditions –
	Carbon oxides, Hydrogen fluoride ¹²
Explosivity (UEL, LEL, shock	No data available
sensitive)	
Oxidizer	No data available
Corrosivity	No data available
рН	No data available
Reactivity	Incompatible materials: Bases, Oxidizing agents, Reducing agents ¹³

Viscosity	No data available
Odor Threshold	No data available
Particle size, shape, respirable	No data available
fraction	
Other physical hazards associated	No data available
with process: Heat, gases under	
pressure, noise, vibration,	
ergonomic hazard	
HEALTH HAZARDS	
Acute Toxicity	
Oral LD ₅₀	"No studies that determined the acute oral LD ₅₀ of pure PFNA were located. However, Mertens et al. (2010) state that the "approximate lethal dose" (unpublished data, calculated herein as 65 mg/kg) in rats for the Surflon S-111 mixture of PFCs consisting primarily of PFNA (see below) was 2.9-fold lower than the acute LD ₅₀ for PFOA of 198 mg/kg identified by Olson and Anderson (1983)." ¹⁴
Dermal LD ₅₀	No data available
Inhalation LC ₅₀	"The inhalation LC ₅₀ in male rats (5 or 6 per group) exposed for 4 hours to six concentrations ranging from 67 to 4,600 mg/m ³ of ammonium perfluorononanoate (the ammonium salt of PFNA) as a dust was 820 mg/m ³ ; the lowest dose that caused death was 590 mg/m ³ . Animals were observed for 5-14 days after exposure and deaths occurred earlier with increasing dose (Kinney et al., 1989). As has been observed in animals acutely exposed to PFOA (reviewed in Lau et al., 2007; Post et al., 2012), severe body weight loss occurred in surviving rats of all but the lowest dose group." ¹⁵
Intraperitoneal LD ₅₀	No data available
Chronic or Sub-chronic Toxicity	
IARC rating	Not found
Carcinogenicity	Not found on the Prop 65 list
Neurotoxicity	Not listed in HAZMAP as a neurotoxin "These results suggested that exposure to PFNA elevated the intracellular calcium level and activated the CaMKII signaling pathway, which may aggravate oxidative stress in PC12 cells and lead to cell damage or cell apoptosis (Fang et al 2018)." ¹⁶
	"A total of 282 subjects have completed the PFASs analysis and questionnaire survey. After adjusted for potential confounders, we observed that PFNA is inversely associated with inattention and oppositional defiant disorder of SNAP-IV, and hyperactivity/inattention of SDQPrenatal exposure to PFNA was found to associate with neurobehavioral symptoms related to ADHD among Asian seven-year-

	old children. Further studies are needed to elucidate the causal relationship (Lien et al 2016)." ¹⁷
	"Serum-PFOS and PFHxS concentrations declined over time, whereas PFOA, PFNA, and PFDA tended to increase. No associations were observed between prenatal PFAS concentrations and SDQ scores. However, a two-fold increase in 5-year serum-PFOA, PFNA, and PFDA concentrations was associated with increases in total SDQ scores by 1.03 (95% CI: 0.11, 1.95), 0.72 (95% CI: 0.07, 1.38) and 0.78 points (95% CI: 0.01, 1.55), respectively. For SDQ subscales, significant associations were found in regard to hyperactivity, peer relationship, and conduct problems, as well as internalizing and externalizing problems and autism screening composite scores. Cross-sectional analyses at age 7 years showed possible sex-dimorphic associations between PFAS concentrations and SDQ scores, where girls had consistently positive associations with SDQ scores whereas boys exhibited a pattern of negative or null associations Higher serum PFAS concentrations at ages 5- and 7-years, but not prenatally, were associated with parent- reported behavioral problems at age 7 (Oulhote et al 2016)." ¹⁸
	"Whereas PFHxS, PFOS and PFUnDA aggregated in large hotspots, PFOA, PFNA and PFDA showed a more dispersed distribution pattern. In conclusion, the toxicity of the investigated PFAAs increased with increasing carbon chain length. For molecules with a similar chain length, a sulfonate functional group led to greater toxicity than a carboxyl group (Berntsen et al 2017)." ¹⁹
Developmental/Reproductive Toxicity	In the EU, Sweden has proposed PFNA to be classified as toxic to reproduction ²⁰
	Not found on the Prop 65 list
	There is evidence based on the RAC opinion on PFNA and its sodium and ammonium salts that these substances meet the criteria for classification as toxic for reproduction in accordance with Article 57 (c) of the REACH Regulation. As a consequence the toxicity criterion of REACH Annex XIII is fulfilled H360Df: May damage the unborn child. Suspected of damaging fertility. ²¹
	ToxPLANET – Notes "Reproductive Effector"
	Oral, rat: TDLo: 70 mg/kg (2W male); Paternal Effects: Spermatogenesis (including genetic material, sperm morphology, motility, and count); Paternal Effects: Other effects on male ²²
	PFNA exposure resulted in significantly altered responses in terms

of morphometric, locomotion, and gene expression endpoints, which could be manifested in field exposed teleosts (Jantzen et al 2016a). ²³
In boys, prenatal PFNA, and PFDoDA were associated with reductions in height at certain ages in childhood, but not with size at birthPrenatal exposure to long-chain PFCAs may interfere with fetal and childhood growth in girls, and childhood growth in boys (Wang et al 2016a). ²⁴
Plasma concentrations ofperfluorononanoic acid (PFNA) were inversely associated with endometriosis-related infertility, but the associations were attenuated in the sensitivity analyses. (Wang et al 2017). ²⁵
We did not find evidence of associations between PFNA and total testosterone or between any of the PFAAs and SHBG Our findings were based on a small study sample and should be interpreted with caution. However, they suggest that prenatal exposure to some PFAAs may alter testosterone concentrations in females (Maisonet et al 2015). ²⁶
In a multiple logistic regression with adjustment for age, BMI, parity and gestational age at serum sampling, women with the highest tertile of exposure to perfluorononanoic acid (PFNA) and perfluorodecanoic acid (PFDA) in pregnancy had odds ratios for miscarriage of 16.5 (95% CI 7.4-36.6-36.5) and 2.67 (1.31-5.44), respectively, as compared to the lowest tertile (Jensen et al 2015). ²⁷ Note: Table 2, on page 6 of Jensen et al 2015, indicates the 95% confidence interval to be 7.39-36.62.
Our results indicate that developmental toxicity of PFNA in mice is comparable to that of PFOS and PFOA, and that these adverse effects are likely common to perfluoroalkyl acids that persist in the body (Das et al 2015). ²⁸
PFNAwere associated with a lower percentage of sperm with coiled tails (Louis et al 2015). ²⁹
The results of this study indicate that chronic exposure to PFNA can lead to dysfunction in the HPGL axis and sex hormone synthesis and cause adverse effects on fish reproduction (Zhang et al 2016d). ³⁰
No significantly elevated HRs were observed for any PFASs suggesting no association with loss:, PFNA (0.86; 0.70, 1.06),, PFOS (0.81; 0.65, 1.00), and PFOA (0.93; 0.75, 1.16) (Louis et al 2016). ³¹
"After adjustment for confounding factors, nonsignificant associations between PFASs and reproductive hormone were found except for PFNA with In(estradiol) (β=0.2060, 95%CI: 0.0016, 0.4105).

When stratified by sex, more significant associations were found in males than in females. Among males, PFASs were negatively associated with ln(testosterone) level for PFNA (β =-0.4233, 95%CI: -0.6998, -0.1467)In conclusion, this study showed higher levels of PFASs coincide with lower testosterone and higher estradiol levels, and more significant associations of PFASs with reproductive hormone were found in males than in females (Zhou et al 2016a)." ³²
"In secondary analyses, instead of PFOS, we examined perfluorononanoate (PFNA) [mean (SD): 0.7 (0.4) ng/mL], a PFAS more closely linked to lower BW/GA in our cohort. The best-fit multi- pollutant model included positive two-way interactions between PFNA and both black carbon and smoking (p-interactions = 0.03) Concurrent prenatal exposures to maternal smoking, black carbon, and PFOS are additively associated with lower fetal growth, whereas PFNA may attenuate associations of smoking and black carbon with lower fetal growth. It is important to examine interactions between multiple exposures in relation to health outcomes, as effects may not always be additive and may shed light on biological pathways (Rokoff et al 2018)." ³³
"The duration of breastfeeding was positively associated with the serum concentrations of In-transformedPFNA(all P<0.001). Height at 2 years of age was inversely related to PFNA, concentrations (adjusted β per In unit [95% confidence interval, CI]: -0.48 [-1.40, -0.51] cm), after adjusting for age, sex, and midparental height. Weight at 2 years of age was inversely associated with PFNA (adjusted β per In unit [95% CI]: -0.32 [-0.48, -0.15] kg), after adjusting for age, sex, and parental BMI. In conclusion, the serum concentrations of PFCs were inversely associated with growth parameters in 2-year-old children (Lee et al 2018)." ³⁴
"Perfluorooctane sulfonate (PFOS) and perfluorononanoate (PFNA) were weakly inversely associated with birth weight-for-gestational age z scores (adjusted β = -0.04 (95% confidence interval (CI): -0.08, 0.01) and adjusted β = -0.06 (95% CI: -0.11, -0.01) per interquartile-range increase, respectively). PFOS and PFNA were also associated with higher odds of preterm birth (e.g., for highest PFOS quartile vs. lowest, adjusted odds ratio = 2.4, 95% CI: 1.3, 4.4). Adjusting for markers of pregnancy hemodynamics (glomerular filtration rate and plasma albumin), to the extent that they accurately reflect underlying pregnancy physiology, did not materially affect associations. These results suggest that pregnancy hemodynamics may not confound

	associations with birth outcomes when PFAS are measured early in pregnancy (Sagiv et al 2018)." ³⁵
	"Inverse associations were found between maternal levels of perfluorononanoic acid (PFNA), perfluorodecanoic acid (PFDA), and perfluoroundecanoic acid (PFUnDA), and birth weight SDS with a change of -0.10 to -0.18 weight SDS for an inter-quartile range (IQR) increase in ng/g PFAA. After birth, weight and length SDS were not significantly associated with maternal PFAA. However, BMI SDS was significantly associated with PFOA, PFNA, and PFHxS at 3 and 4years of age, and with PFOS at 4 and 5years of age. If causal, these associations suggest that PFAA affects fetal and childhood body development in different directions (Gyllenhammar et al 2018)." ³⁶
	"Median serum PFASs concentrations were PFOS>PFOA>PFHxS>PFNA in prenatal, 3-year, and 8-year children. The covariate-adjusted general linear regression identified positive associations between serum PFOA, PFOS and PFNA concentrations and children's reading scores at ages 5 and 8 years, Prenatal PFOS and PFNA were positively associated with children's reading abilities at age 5 years, but not at age 8 years; 3- year PFOS and PFNA were positively associated with reading scores at age 5 years Prenatal and childhood serum PFOA, PFOS and PFNA concentrations were positively associated with better children's reading skills at ages 5 and 8 years, but no association was found between serum PFHxS and reading skills (Zhang et al 2018)." ³⁷
Genotoxicity/Mutagenicity	We concluded that PFOA and PFNA induced DNA damage and the biomarker of oxidative DNA damage (8- OHdG) could be measured by HPLC-MS/MS. In addition, PFNA produced high level of 8-OHdG at concentrations lower than PFOA, this may indicate that PFNA is more potent genotoxicant for TK6 cells than PFOA (Yahia et al 2016). ³⁸ "This is the first report of the effects of perfluorinated acids on the activity of purified enzymes. The results show these compounds have a very low acute biological activity. The observed effective concentrations lie in the millimole range, which is well above probable intracellular concentrations. A relationship was found between the toxicity of the perfluorinated carboxylic acids and the perfluorocarbon chain length: in every test system applied, the longer the perfluorocarbon chain, the more toxic was the acid. The lowest effective concentrations were thus recorded for perfluorononanoic and perfluorodecanoic acids (Mulkiewicz et al 2007)." ³⁹

	values of EC_{50} decreased with elongation of fluorocarbon chain (PFHxA
	> PFHpA > PFOA > PFNA > PFDA > PFDoA > PFTeDA). Further elongation
	(C16 and C18) did not deepen the effect but even partially reversed it.
	The effect was intensified after longer exposure (72 h); (Kleszczyński
	<mark>2007</mark>)." ⁴⁰
Endocrine Disruption	Found on TEDX List of Potential Endocrine Disruptors ⁴¹
	Maternal serum concentrations of HCB, PFOS and PFOA were
	associated with increased BMI z-scores and/or overweight risk (i.e. BMI
	z-score≥85th WHO percentile). No clear association was found for
	maternal serum-PCBs, p,p'- DDE, PFHxS, PFNA and PFDA. (Karlsen et al.,
	<mark>2017</mark>). ⁴²
	Overall, this study demonstrated that PFAA mixture could have the
	potential of multigenerational endocrine disruptors in O. latipes (Lee et
	al 2017). ⁴³
	Normal acini maturation was negatively impacted by PFOS, PFNA and
	PFDA already at the lowest concentration tested (0.6μM). Observed
	effects included loss of organization of the cell clusters and absence of
	a hollow lumen. Overall, this study demonstrated that PFAAs can
	interfere with cellular events related to normal development of
	glandular breast tissue through ER-independent mechanisms (Halsne et
	al 2016). ⁴⁴
Thyroid	Total PFAS exposure level was 2.63-44.7ng/mL in the case group and
	2.44- 22.4ng/mL in the control group. Concentrations of
	serumperfluorononanoic acid (PFNA, p<0.001), were significantly
	higher in the case group than the control group (Kim et al 2016).45
	"We also highlighted a negative association between
	perfluorononanoic acid (PFNA) concentration and TSH in male
	newborns (p = 0.018) ($\frac{1}{\text{Dufour et al 2018}}$)". ⁴⁶
	"PFAS concentrations [PFOA, PFOS, and perfluorononanoate (PFNA)]
	were inversely associated with TSH levels in TPOAb-positive women
	only (Preston et al 2018)". ⁴⁷
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	"We found some evidence of a positive association between PFNA
	and TSH levels measured in the blood of boys aged ≥11 years.
	Although there is a small number of studies with comparable data,
	we found some consistency of a positive association between maternal
	or teenage male exposure to some PFAS and TSH levels based on the
	current literature. However, further studies are required to confirm
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these possible relationships (Ballesteros et al 2017)."48 "PFAS levels gradually increased from 2006 to 2013, decreasing thereafter. We found that PFAS levels were higher in male than in female participants and were positively correlated with age. PFASs were not significantly correlated with body mass index, although we observed positive correlations with total cholesterol, low-density lipoprotein cholesterol, and triglycerides and negative correlations with high-density lipoprotein cholesterol. Uric acid and free thyroxine (fT4) also showed positive correlations with major congeners while correlations between thyroid stimulating hormone and PFASs were inconsistent. We demonstrated significant correlations between fT4 and perfluorononanoic acid (PFNA), ... Furthermore, principal component analysis suggested possible differences in disease manifestation based on the congener distribution of PFASs (Seo et al 2018)."⁴⁹ Hormone Summary – Females Free thyroxine is decreasing with increasing dose. Hormone Summary – Males Free Thyroxine (T4) and TSH are decreasing with increasing dose. See additional information in NTP Tox-97 report.⁵⁰ Immunotoxicity The substance is immunotoxic (Fang et al., 2008)⁵¹ "In ovalbumin (OVA)-induced model of systemic anaphylaxis in the presence of hypothermia, PFNA, PFDA, and PFUnA exacerbated allergic symptoms accompanied by elevation in serum histamine, TNF- α , IgE, and IgG1. Our data indicate that some PFC aggravated high-affinity IgE receptor (FceRI)-mediated mast cell degranulation and allergic symptoms. Consequently, the results demonstrated that carbon-chain length of PFC may serve as a factor in allergic inflammation (Lee and Kim 2018)".52 "The number of reported airways infections were significantly associated with cord blood concentrations of PFAS; ...LRTIs from 0 to 10 years of age with PFOS (β = 0.50 (0.42-0.57)), PFOA (β = 0.28 (0.22-0.35)), PFOSA (β = 0.10 (0.06-0.14)), PFNA (β = 0.09 (0.03-0.14)) and PFUnDA (β = 0.18 (0.13-0.23)) concentrations. Neither reduced lung function at birth, asthma, allergic rhinitis, AD nor allergic sensitization were significantly associated with any of the PFASs. ... Although prenatal exposure to PFASs was not associated with atopic or lung manifestations by 10 years of age, several PFASs were associated with an increased number of respiratory tract infections in the first 10 years

of life, suggesting immunosuppressive effects of PFASs (Impinen et al 2018)."53

"...A total of 687 children completed a 2-year follow-up visit and had PFASs measurement. AD (atopic dermatitis) was diagnosed in 173 (25.2%) children during the first 24 months. In female children, a logunit increase in perfluorooctanoic acid (PFOA) was associated with a 2.1-fold increase in AD risk (AOR 2.07, 95% CI 1.13-3.80) after adjusting for potential confounders. The corresponding risk was 2.22 (1.07-4.58) for perfluorononanoic acid (PFNA). ... The highest quartile of PFNA, perfluorodecanoic acid (PFDA) and perfluorohexane sulfonic acid (PFHxS) were associated with AD with AOR (95% CI) being 2.14 (0.97-4.74), 2.14 (1.00-4.57), and 2.30 (1.03-5.15), respectively. ... However, no significant associations were found in male children. ... In addition, the associations between AD with prenatal exposure to PFNA were close to statistical significance (Chen et al 2018)."⁵⁴

"Together, these results suggest that PFNA exerts toxic effects on lymphoid organs and T cell and innate immune cell homeostasis in mice and that these effects may result from the activation of PPAR-alpha, PPAR-gamma, and the hypothalamic-pituitary-adrenal axis. Interestingly, at the transcriptional level, the nuclear factor-kappa B signaling pathway appears to be uninvolved in the immunotoxic potential of PFNA (Fang et al 2008)."⁵⁵

"The present study demonstrates that PFNA produces immunotoxic effects in both male and female C57BL/6 mice as evidenced by splenic atrophy, decreased splenocyte numbers, and a marked reduction in thymocyte viability. The current study also demonstrates that the effects of PFNA on different leukocyte populations are not uniform. The CD4⁺CD8⁺ double-positive thymocytes were particularly sensitive to PFNA in which the proportion of this population was >95% decreased relative to the entire CD4⁺ thymocyte population in PFNA-treated mice. Interestingly, PFNA also markedly increased serum levels of TNF α in response to LPS in mice. Collectively, the present studies demonstrate that PFNA decreases lymphocyte viability and alters the immune response to LPS in C57BL/6 mice (Rockwell et al 2013)."⁵⁶

"The present studies sought to determine whether, and to what degree, the immune system recovered 28 days after PFNA exposure. None of the parameters measured had fully recovered. A few parameters had partially recovered, including decreased spleen size

	and the decreased ratio of the CD4 ⁺ /CD8 ⁺ double-positive population in thymus. The majority of effects of PFNA remained unchanged 28 days after exposure, including decreased proportion of intact thymocytes (as determined by FSC vs SSC), alterations in the ratios of immune cell populations in spleen and the CD4 ⁺ , CD8 ⁺ and double-negative populations in thymus. Notably, PFNA markedly increased the TNFα response to LPS in vivo, and no recovery was evident 28 days after exposure. The effect of PFNA on CD4 ⁺ T cells, CD8 ⁺ T cells and CD19 ⁺ cells was more pronounced in females. The current study demonstrates that a single high dose exposure to PFNA (e.g. as might occur accidentally in an occupational setting) has long-lasting effects on the immune system (Rockwell et al 2017)." ⁵⁷
	"Thymic and/or splenic alterations were observed in rats and mice administered ≥1 mg/kg/day PFNA (Fang et al. 2008, 2009, 2010)". ⁵⁸
	Also see Liu & Gin 2018 in the Ecological Toxicity section below
Heart	"When adjusting for multiple comparisons, none of the eight PFASs evaluated were significantly related to left ventricular mass. However, perfluorononanoic acid (PFNA), perfluorodecanoic acid (PFDA), and perfluoroundecanoic acid (PFUnDA) were related to relative wall thickness (RWT) in a negative fashion (p < 0.0021). Besides being inversely related to RWT, PFNA was also positively related to left ventricular end-diastolic volume (LVEDD) (p < 0.0021)In this cross- sectional study, several of the PFASs evaluated, especially PFNA, were related to myocardial geometry: a reduction in relative wall thickness and an increase in left ventricular diameter following adjustment for traditional cardiovascular risk factors, suggesting a role for PFASs in cardiac remodeling (Mobacke et al 2018)." ⁵⁹
Liver/Lipid Effects	Increases in liver weight and cell size, and decreases in DNA content per mg of liver, were observed for all compounds in WT mice, and were also seen in PPAR α -null mice forPFNA, (Das et al 2017). ⁶⁰
	These results indicate that most of the PFAAs increase liver TG load and promote steatosis in mice (Das et al 2017). ⁶¹
	ITC measurement revealed that PFOA/PFNA displayed a moderate affinity for hL-FABP at a 1:1 molar ratio, a weak binding affinity for PFHxS and no binding for PFHxA (Sheng et al 2016). ⁶²
	In summary, all perfluorinated compounds increased cell number, decreased cell size, increased total triglyceride, and altered expression

	of genes associated with adipocyte differentiation and lipid metabolism (<mark>Watkins et al, 2015</mark>). ⁶³
	"Moreover, PFNA increased free and total cholesterol in mouse liver but not in mouse serum. Furthermore, PFNA increased mRNA expression of sterol transporters, namely Abca1, g1, g5/g8, and steroidogenic acute regulatory protein via PPARα. In conclusion, PFNA produced cholestasis in mouse liver, and the activation of PPARα plays a central role in regulating BA and cholesterol metabolism and transport in mouse serum and liver (Zhang et al 2018a)." ⁶⁴
	"While PFASs exposure was associated with a consistent negative relationship with bone health parameters, among four PFASs tested, only PFNA showed a significant negative relationship with bone parameter (β [95% CI], = -72.7 [- 126.0, - 19.6], p = .010). PFNA was also associated with raised systolic blood pressure (p = .008), low density lipoprotein cholesterol (LDL-C; p < .001), and total cholesterol (TC; p = .014) In this analysis, PFASs were not strongly related to thyroid hormones, 25-hydroxy vitamin D, liver enzymes, or glucose homeostasisPFASs exposure in obese children may play a role in adverse skeletal and cardiovascular risk profiles (Khalil et al 2018)." ⁶⁵
	"Perfluorohexanoate, perfluoroheptanoate, perfluorooctanoate, perfluorononanoate (PFNA) and perfluorodecanoate induced PPARα activities >2.5-fold compared to controls. The concentration-response relationships were positive for all the tested compounds, except perfluorooctane sulfonate PFOS and FOSA, and were compound- specific, as demonstrated by differences in the estimated slopes. The relationships were steeper for PFCAs with chain lengths up to and including PFNA than for the other studied PFASs (Rosenmai et al 2017)." ⁶⁶
Diabetes/Obesity	"The expression of phospho-glycogen synthase kinase-3 beta (GSK3β, Ser 9) was increased, which explains the augment of hepatic glycogen. Significant increases in hydrogen peroxide (H(2)O(2)) and malondialdehyde (MDA) were found in the livers of 5mg/kg/d PFNA- treated rats. Thus, exposure to PFNA disordered glucose metabolism via inhibiting hepatic insulin signal pathway, accelerating the output of glucose and increasing glycogen synthesis in the rat liver. Furthermore, the oxidative stress induced by PFNA may be involved in this process (Fang et al 2012)". ⁶⁷
	The expressions of protein related to lipid homeostasis, liver X receptor

	 α and apolipoprotein E, were decreased after PFNA administration. Exposure to PFNA also increased the activity of serum alanine aminotransferase in diabetic rats this study discloses that exposure to PFNA impacts on enzymes and proteins related to liver lipid metabolism and lead to obvious accumulation of lipid in the liver of diabetic rats, which may be responsible for hepatotoxicity of this compound in individuals with diabetes mellitus (Fang et al, 2015).⁶⁸ However, PFNA, showed a potential protective effect against glucose intolerance and the risk of diabetes (Su et al 2016).⁶⁹ "PFHxS and PFNA concentrations were associated with impaired glycemic status in metabolically vulnerable pregnant women and might further enhance the risk of developing GDM (gestational diabetes mellitus)(Jensen et al 2018)."⁷⁰ "Furthermore, diabetes was not related to PFOS, PFHxS and PFNA, regardless of gender (He et al 2018)."⁷¹ "In women, comparing the highest to the lowest tertiles of PFAS concentrations, the multivariate-adjusted mean weight regain (SE) was 4.7 (0.9) versus 2.5 (0.9) kg for perfluorononanoic acid (PFNA) (Ptrend = 0.006); When further adjusted for changes in body weight or thyroid hormones during the first 6 months, results remained similar. Moreover, higher baseline plasma PFAS concentrations, especially for PFOS and PFNA, were significantly associated with greater decline in RMR during the weight-loss period and less increase in RMR during the weight regain period in both men and women (Liu et al 2018).
	al 2018)." ⁷²
Skin, Eye and Respiratory Effects	
Irritant – Skin, Eye, or Respiratory	Skin Irrit. 2; Eye Irrit. 2A; STOT SE 3; H315, H319, H335 ⁷³
Corrosive – S, E, or R	Corrosive as concentrated acid; pKa = -0.21 ⁷⁴
Permanent Damage – S, E, or R	
Sensitizer-S&R	
Asthmagen – Initiator or	Not found in AOEC Database as of 01/04/18
Exacerbator	
Skin Absorption, Kp	
LOAEL	12,400 ng/ml – pregnant mice (<mark>Das et al., 2015</mark>); 11,500 ng/ml – male mice (Wang et al, 2015) ⁷⁵
NOAEL	
Benchmark Dose Response (BMD)	Benchmark Dose Modeling for 10% Increase in Liver Weight in Pregnant Mice from Das et. Al (2015) – Table 10 in NJ 2015 document ⁷⁶
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Toxicokinetics	a. PFNA is present in human blood of the general population
	b. Elimination half-lives are > 1.7 years
	c. Human elimination half-lives seem to be the longest amongst the
	available mammalian data, whereas the elimination half-lives in
	laboratory mammals vary highly depending on the study conditions
	(<mark>ECHA 2015</mark>). ⁷⁷
Metabolites	
Synergistic or Antagonistic Effects	
Environmental and Human Health	Exposure and Risk Values
RfC/RfD	Not found in IRIS as of 12/14/17
ATSDR-MRL	Not found on June 2017 MRL list as of 01/04/18; Provisional Oral
	Minimal Risk Level (MRL) – Intermediate – 3x10 ⁻⁶ mg/kg/day. Most
	sensitive targets were body weight and developmental endpoints (Das
	et al 2015). ⁷⁸
Adverse Effect Levels: DNEL, PNEC,	
PNEL	
Health Based Exposure Limits	
NIOSH-REL/IDLH/Ceiling Limits	Not found
OSHA-PEL	Not found
ACGIH TLV-TWA	Not found
TLV-STEL	Not found
Biomonitoring Action Limits	Connecticut DPH set a drinking water Action Level for private wells in
	2016 for PFAS that is the same as the EPA Health Advisory (70 ppt) but
	has added three additional PFAS (PFNA, PFHxS, PFHpA) to the group.
	The sum of this group of 5 PFAS must be below the target
	concentration of 70 ppt. These additional PFAS have produced some of
	the same health effects as PFOS and PFOA (CT DPH 2017). ⁷⁹
Drinking Water Standards	NJ: Drinking Water Quality Institute recommends that the Department
	propose and adopt an MCL of 13 ng/L for PFNA in drinking water ⁸⁰
	based on a study of developmental effects in which pregnant mice
	were exposed to PFNA for 16 days (Das 2015). The Health-based MCL is
	further supported by data on effects in the offspring in the same study,
	and on increased liver weight and other effects in additional rodent
	studies from the same and other laboratories. ⁸¹
Other	Mechanism of Action information in HSDB
	On SIN List (reason for inclusion): This substance has persistent,
	bioaccumulative and toxic properties, has been detected in human
	blood, in dolphins, seals and polar bears. Substance is concluded to be
	a PBT and Reprotoxic substance SVHC by ECHA Member State
	Committee. ⁸²
ENVIRONMENTAL & ECO-SYSTEM L	IAZARDS (Bulk of information cited ECHA 2015)
PBT/Persistence	Persistence : "PFNA is, based on its stabile structure, not expected to
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	undergo abiotic degradation under relevant environmental
	conditions. ^{"83}
	Biodegradation in water: "half-life in water of 2477 days and a half-
	life in soil of 4954 days were estimated (Lambert et al., 2011).
	Nevertheless, these estimates are assumed to be low because the bond
	between carbon and fluorine is one of the most stable ones in organic
	chemistry and not subject to degradation by microorganisms occurring
	in the environment." ⁸⁴
	Screening test for ready biodegradability: no biodegradation was
	observed after 28 days (Stasinakis et al., 2008) ⁸⁵
	No photodegradation tested under relevant environmental conditions;
	100 % after 12 h (irradiated with a xenon-mercury lamp under oxygen)
	by use of persulfate ion $(S_2O_8^{2-})$ in water (Hori 2005).
	No hydrolysis tested under relevant environmental conditions; 100%
	(by use of S2O82-) after 6 h in 80°C water. (<mark>Hori 2008</mark>) Reaction
	products were mainly F ⁻ and CO ₂ ; short chain PFCAs were minor
	reaction products (ECHA 2015). ⁸⁶
	Presence in the environment:
	"PFOS, PFOA, PFHxS, PFHpA [Median concentration in dust: 10 – 97.3
	ng/g] and PFNA were found in all of the vacuum dust samples and dust
	from eight homes contained all 16 PFCs included in our analysis
Bioaccumulation	(Knobeloch et al 2012). ^{"87} "Due to its expected notable water solubility, PFNA is, like the other
Diodecumulation	PFCAs, expected to quickly be excreted via gill permeation.
	Furthermore, PFNA is present mainly in protein rich tissues like blood
	and liver (OECD, 2006; Kelly et al. 2009). Hence, bioconcentration in gill
	breathing organisms and the accumulation in lipids is not the most
	relevant endpoint to consider. Field studies show that air-breathing
	organisms are more likely to bioaccumulate PFNA and other PFCAs
	compared to water breathing organisms. Therefore, the numerical
	bioaccumulation (B) criterion defined in the REACH regulation Annex
	XIII (sections 1.1.2 and 3.2.2(a)) is not suitable to assess the
	bioaccumulation potential of PFNA (ECHA 2015)." ⁸⁸
	The bioaccumulation data on PFNA in environmental species, in
	laboratory mammals and in humans is consistent with the data on
	other long-chain perfluorinated carboxylic acids, such as PFOA. a.

Recent models to explain the substantial bioaccumulation of PFCAs take into account the observed pattern of animal tissue distribution, the relationship between chain length and bioaccumulation and the species and gender-specific variation in elimination half-life. ⁸⁹ To conclude, taken all available information together in a weight-of-evidence approach, the elimination half-lives from humans and other mammals show that PFNA bioaccumulates. The available field data also indicate that bioaccumulation and trophic magnification occur in certain food webs in the environment. The data on PFNA are in line with the expected regular pattern of fate properties of the already assessed PFOA and C11-C14-PFCAs. Therefore it is considered that the B criterion of REACH Annex XIII is fulfilled. Whether the vB criterion is fulfilled has not been assessed. ⁹⁰
 breathing mammals, including endangered species and humans a. BMFs range from 1.4 – 24 based on estimated whole body values b. TMFs range from 2.9 to 9.88 referring to either whole body measurements or estimated whole body values
 3. PFNA does not seem to consistently accumulate in water breathing animals a. No experimental BCFs are available for PFNA. For the closest structural analogues BCFs range from 4.0 to 27 (PFOA) and from 450 to 2,700 (PFDA) b. Whole body BAFs range from 0 to 3,981 c. Whole body BMFs range from 0.13 to 5.3 whereas most of the data are below 1
 d. Whole body TMFs range from [0.33 to 2.1] in aquatic piscivorous food webs⁹¹ Presence in air breathing mammals: PFNA was identified in female beaver tissues (liver, subcutaneous
adipose and peritoneum) in a range from 1.50 to 6.61ngg(-1) ww (<mark>Surma et al, 2015</mark>). ⁹²
Presence in marine mammals: "Juveniles (dolphins) had significantly higher levels of PFOS and PFNA

than adults in both species, suggesting growth dilution as they approach maturity (Lynch et al 2018)."93 Presence in humans: The geometric mean concentrations (and P95 values) for ... PFNA (perfluorononanoic acid) in the serum of 755 Spanish adults, 0.96 (2.44) µg/L, (Bartolome et al 2017).⁹⁴ "...perfluorononanoate (PFNA) ... found in 99% of the participants. ... Median concentrations were: ... PFNA 0.50 ng/mL.... Median of PFASs sum concentration (Σ PFAS) was 10.7 ng/mL, the concentration range was 2.6-200.8 ng/mL. Intake of fat fish, fish liver, seagull eggs, reindeer meat and drinks with sugar were the main dietary predictors of several PFASs. Intake of junk food (pizza, hamburger, sausages) was positively associated with PFNA, intake of canned food was positively associated with PFHxS (Averina et al 2018)."95 "We detected PFOS, PFOA, PFHxS, and PFNA in all children at concentrations similar to those of NHANES 2013-2014 adolescents and adults, suggesting prevalent exposure to these PFAS or their precursors among U.S. 3-11 year old children, most of whom were born after the phase out of PFOS in the United States in 2002 (<mark>Ye et al 2018</mark>)."⁹⁶ "After a month of nursing, the concentrations of PFOS, PFOA, perfluorononanoic acid (PFNA), and ΣPFAS significantly increased. This could be due to changes in the dietary and behavior patterns of the mothers after the first month of lactation. ... Certain types of diet (e.g. consuming snacks and milk) and eating-out frequency were significantly associated with increasing levels of PFAS (Lee et al 2018a)."⁹⁷ "Serum levels of perfluorohexane sulfonic acid (PFHxS), perfluorooctanoic acid (PFOA) and perfluorononanoic acid (PFNA) were significantly higher in the subjects with MetS (metabolic syndrome). Logistic regression results showed that concentration of PFNA in serum was associated with 10.9-fold [95% confidence interval (CI), 2.00-59.1]

increased risk of MetS. Moreover, increased serum PFNA

	concentrations were associated with high blood pressure [both for systolic and diastolic blood pressure (SBP and DBP); odds ratio (OR) 7.52 (95%Cl, 1.34-42.1) for SBP and 7.27 (95%Cl, 1.17-45.1) for DBP], hypertriglyceridemia [13.2 (95%Cl, 2.34-74.2)] and obesity [13.3 (95%Cl, 2.38-74.4)], respectively (Yang et al 2018)." ⁹⁸
	In an NHANES 2013-2014 analysis: "Diet accounted for a low of 18.6%
	of the total explained variance in the adjusted levels of NPFOA (linear
	isomer of PFOA) and a high of 72.3% for PFNA (<mark>Jain 2018</mark>)." ⁹⁹
	Phytoaccumulation – PFHpA, PFOA, PFNA and PFBS were taken up by
	yam root, maize cob and sugarcane stem (Dalahmeh et al 2018). ¹⁰⁰
BAF	BAFs range from 0 to 3981 ¹⁰¹
	BAFs: [biota]/[water]
	Phytoplankton (whole): 1680 (<mark>Loi et al, 2011</mark>)
	Lake trout (whole): 3981 (Furdui et al, 2007)
	European chub (various tissues): 39 (muscle) – 630 (plasma) (<mark>Labdie</mark>
	and Chevreuil, 2011)
	Atlantic croaker (whole): not detected (Houde et al 2006)
	5 different marine fish (whole): 705-2800 (<mark>Houde et al 2006</mark>)
	5 different marine fish (whole) 59-197 (<mark>Loi et al, 2011</mark>)
	Biota-sediment accumulation factor (BSAF) for benthic-dwelling worm
	downstream from 2 WWTP in California varied from 83-149 (lipid
	normalized) 0.64 – 0.83 (non-lipid normalized). Use results with
	caution, as PFNA does not enrich in lipids. (Higgins et al, 2007) ¹⁰²
BCF	PFNA does not seem to consistently accumulate in water breathing animals; No experimental BCFs are available for PFNA. For the closest structural analogues BCFs range from 4.0 to 27 (PFOA) and from 450 to 2700 (PFDA) ¹⁰³
BMF	There is evidence that PFNA preferentially bioaccumulates in air-
	breathing mammals, including endangered species and humans: BMFs
	range from 1.4 – 24 based on estimated whole body values 104
	(referencing <mark>Houde 2006</mark> dolphin-prey study)
	Beluga whale liver BMFs with various prey ranged from 1.2 – 12.9
	(Tomy 2009) ¹⁰⁵

,,	Gill/water breathing animals: whole body BMFs range from 0.13 to 5.3
	whereas most of the data are below 1^{106}
Ecological Toxicity	PFNA was also the compound producing the greatest levels of oxidative stress, both in zebrafish and TLT cells. Additionally, in both biological systems, it showed a much stronger effect on mixtures in comparison to the others PFCs tested in this work (Rainieri et al 2017). ¹⁰⁷
	This study demonstrates that acute, embryonic exposure (5 days) to individual PFASs result in significant biochemical and behavioral changes in young adult zebrafish 6 months after exposure. These three PFASs have long term and persistent impacts following short term embryonic exposure that persists into adulthood (Jantzen et al 2016). ¹⁰⁸
	The observed dose-response PFAS-induced effects were to some extent related to their cytotoxicity: the EC ₅₀ -values of most influential PFAS treatments increased (PFOS < PFNA < PFOA \ll PFBS), and higher-doses of these chemicals induced a larger impact. Major spectral alterations were mainly attributed to DNA/RNA, secondary protein structure, lipids, and fatty acids. Finally, PFOS and PFOA caused a decrease in A6 cell numbers compared to controls, whereas PFBS and PFNA did not significantly change cell population levels. Overall, this work highlights the ability of PFASs to alter A6 cells, whether forming monolayers or differentiated into dome structures, and the potential of PFOS and PFOA to induce cell death (Gorrochategui et al 2016). ¹⁰⁹
	All three PFCs commonly resulted in a decrease in total body length, increased tfc3a (muscle development) expression and decreased ap1s (protein transport) expression at 5dpf, and hyperactive locomotor activity 14 dpf. All other endpoints measured at both life-stage time points varied between each of the PFCs. PFOS, PFNA, and PFOA exposure resulted in significantly altered responses in terms of morphometric, locomotion, and gene expression endpoints, which could be manifested in field exposed teleosts (Jantzen et al 2016a). ¹¹⁰
	Acute toxicity was greater for PFOS in zebrafish; however, it was greater for PFNA in TLT cells. PFNA was also the compound producing

	the greatest levels of oxidative stress, both in zebrafish and TLT cells.
	Additionally, in both biological systems, it showed a much stronger
	effect on mixtures in comparison to the others PFCs tested in this work.
	Mixture studies in zebrafish showed that acute toxicity depended on
	the concentration and that the mixture was far more toxic than the
	individual compounds (<mark>Rainieri et al 2017</mark>). ¹¹¹
	,
	These findings suggest that PFNA affected the development of zebrafish embryos at relatively low concentrations (Liu et al, 2015). ¹¹²
	Immunotoxicity in green mussels: "We found that exposure to PFASs could lead to reduced hemocyte cell viability and suppress immune function by up to 50% of normal performance within the experimental exposure range. The results indicate that PFASs have an immunotoxic potential and thus could pose severe health risks to aquatic organisms. The reported immunotoxicity is likely to result from the compounds' direct and indirect interactions with the hemocyte membrane, and therefore likely to affect the functionality of these cells. The immunotoxic response was found to be related to the organism's burden of PFASs, and was reversible when the compounds were removed from the test organisms (Liu & Gin 2018)." ¹¹³
	"Earthworms were exposed to spiked soil for 21 days. Concentrations of these compounds in earthworms after 21-d exposure ranged from below detection to 127 mg kg-1 wet weight with the rank order of PFNA > PFHxS > PFHpA > PFBS; no mortality of earthworms was observed in all treatments including controls, except PFBS at 1,000 μg kg-1 and all PFASs at 100,000 μg kg-1. The highest weight loss (29%) was observed for earthworms exposed to PFNA at 100,000 μg kg-
	1, which was significantly different from all other treatments except
	PFHpA at 100,000 μ g kg-1 (Karnjanapiboonwong et al 2018). ^{"114}
Aquatic Toxicity: LC ₅₀ , EC ₅₀ , ErC ₅₀ ,	μ
NOAEC/NOEC	
Mammalian Toxicity: LC50, EC50,	
ErC ₅₀ , NOAEC/NOEC	
Wildlife Toxicity: LC ₅₀ , EC ₅₀ , ErC ₅₀ ,	
NOAEC/NOEC	
Breakdown/degradation	
/combustion products	
Anaerobic degradation	

Aerobic degradation	
-	Study examining interaction of DEAS with linid call membrane, to
Other observable ecological	Study examining interaction of PFAS with lipid cell membrane, to
effects (e.g. BOD)	advance understanding of bioaccumulative and toxic potential: "All
	PFASs were found to interact with the [phospholipid] bilayer by
	incorporation, indicating PFAS ability to accumulate once ingested or
	taken up by organisms. The interactions were observed to increase
	with chain length and vary with the functional group as
	SO ₂ NH ₂ (FOSA)>SO ₂ O-(PFOS)>COO-(PFNA). The PFAS hydrophobicity,
	which is strongly correlated with perfluorocarbon chain length, was
	found to strongly influence the interactions. Longer chain PFASs
	showed higher tendency to penetrate into the bilayer compared to the
	showed higher tendency to perfect the find the bindyer compared to the showed higher tendency to perfect the find the bindyer compared to the showed higher tendency to perfect the find the bindyer compared to the bindyer c
	substances but one (PFNA) be removed from the lipid membrane by
	gentle rinsing with water (2mLmin-1). Although short-chain PFASs have
	been suggested to be the potentially less bioaccumulative alternative,
	we found that in high enough concentrations they can also disturb the
	bilayer (<mark>Nouhi et al 2018</mark>)." ¹¹⁵
Fate and Transport: Aquatic	
Fate and Transport: Terrestrial	
Fate and Transport: Atmospheric	
Transport Issues	
Factors affecting bioavailability	
Global Environmental Impacts	
Ozone Depletion Potential (ODP)	
Global Climate Change	
Greenhouse Gas Production	
Acid Rain Formation	
Special Reports	
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